



## **Status and Trends of Prey Fish Populations in Lake Michigan, 2004<sup>1</sup>**

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### Abstract

The Great Lakes Science Center (GLSC) has conducted lake-wide surveys of the fish community in Lake Michigan each fall since 1973 using standard 12-m bottom trawls towed along contour at depths of 9 to 110 m at each of seven to nine index transects. The resulting data on relative abundance, size structure, and condition of individual fishes are used to estimate various population parameters that are in turn used by state and tribal agencies in managing Lake Michigan fish stocks. All seven established index transects of the lake-wide survey were completed in 2004. Alewife abundance decreased substantially between 2003 and 2004, as estimated lake-wide biomass dropped from 43.234 kilotonnes (kt) (1 kt = 1000 metric tons) in 2003 to 13.721 kt in 2004. Catch of adult alewives in 2004 was dominated by the 1998 year-class. Lake-wide biomasses of bloater, deepwater sculpin, slimy sculpin, and rainbow smelt in 2004 were estimated at 34.321 kt, 26.519 kt, 3.658 kt, and 1.854 kt, respectively. Bloater biomass drastically declined between 1989 and 2004. Abundance of juvenile bloaters in 2004 was the highest juvenile bloater abundance observed since 1991. Rainbow smelt biomass declined during 1992-1997, and has remained low since 1997. Deepwater sculpin biomass has shown neither an increasing nor decreasing trend from 1990 to 2004. Burbot abundance decreased during 2001-2003, but increased in 2004. Slimy sculpin abundance appeared to be leveling off during 1999-2004, following an increase during the 1990s. Yellow perch year-class strength was poor in 2004. Lake-wide biomass estimates of dreissenid mussels trended neither upward nor downward during 1999-2004. Total catch of round gobies in 2004 (26 fish) was similar to that in 2003 (23 fish), and catches were again limited to the Manistique and Saugatuck transects.

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The Great Lakes Science Center (GLSC) has conducted daytime bottom trawl surveys in Lake Michigan during the fall annually since 1973. From these surveys, the relative abundance of the prey fish populations are measured, and estimates of lake-wide biomass available to the bottom trawls (for the region of the main basin between the 5-m and 114-m depth contours) can be generated (Hatch et al. 1981; Brown and Stedman 1995). Such estimates are critical to fisheries managers making decisions on stocking and harvest rates of salmonines and allowable harvests of fish by commercial fishing operations.

The basic unit of sampling in our surveys is a 10-minute tow using a bottom trawl (12-m headrope) dragged on contour at 9-m (5 fathom) depth increments. At most survey locations, tows range from 9 or 18 m to 110 m. Age determinations are performed on alewives and bloaters from our bottom trawl catches (TeWinkel et al. 2002; Madenjian et al. 2003). Although our surveys have included as many as nine index transects in any given year, we have consistently conducted the surveys at seven transects. These index transects are situated off Manistique, Frankfort, Ludington, and Saugatuck, Michigan; Waukegan, Illinois; and Port Washington and Sturgeon Bay, Wisconsin (Figure 1). All seven index transects were completed in 2004.

Lake-wide estimates of fish biomass require (1) accurate measures of the surface areas that represent the depths sampled and (2) reliable measures of bottom area swept by the trawl. A complete Geographical Information System (GIS) based on depth soundings at 2-km intervals in Lake Michigan was developed as part of the acoustics study performed by Argyle et al. (1998). This GIS database was used to estimate the surface area for each individual depth zone surveyed by the bottom trawls. For 2004, we obtained a more accurate estimate of surface area in the 46-m depth zone, which will slightly modify our historical lake-wide biomass estimates. Trawl mensuration gear that monitored net configuration during deployment revealed fishing depth ( $D$ , in meters) to influence the bottom area swept by the trawl. Since 1998, we have corrected the width ( $W$ , in meters) of the area sampled according to  $W = 9.693 - 43.93/D$ , as well as the actual time ( $AT$ , in minutes) spent on the bottom according to  $AT = \text{tow time} - 3.875$

+  $D^{0.412}$  (Fleischer et al. 1999). These relationships, along with boat speed, were used to estimate bottom area swept.

Beginning this year, we have made several changes in our reporting of the bottom trawl survey data. To better facilitate comparison of our estimates of fish abundance in Lake Michigan with abundance estimates in other lakes and with hydroacoustic estimates of abundance, we have opted to report fish density, expressed both as number of fish per hectare (ha) and kg of fish per ha, rather than catch per tow. A weighted mean fish density over the entire range of depths sampled (within the 5-m to 114-m depth contours) was estimated by first calculating mean density for each of the depth zones, then weighting each depth zone mean density by the respective proportion of lake surface area assigned that depth zone. Beginning this year, estimates of variability in estimates of mean fish density and lake-wide biomass are also reported. Standard error (SE) of mean fish density was estimated by weighting the variances of fish density in each of the depth zones by the appropriate weight (squared proportion of surface area in the depth zone), averaging the weighted variances over all depth zones, and taking the square root of the result.

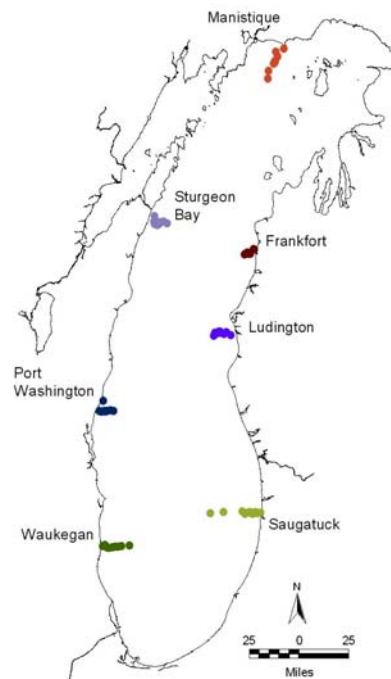


Figure 1. Established sampling locations for GLSC bottom trawls in Lake Michigan.

Relative standard error (RSE) was calculated by dividing SE by mean fish density and then multiplying this ratio by 100 to yield a percentage. SE and RSE for the estimate of lake-wide biomass were calculated in a manner analogous to that for calculating SE and RSE for the estimate of mean fish density.

### ABUNDANCE

By convention, we classify "adult" prey fish as those individuals age 1 or older. Life stage classification was assigned based on length-frequency, where alewives greater  $\geq 100$  mm, rainbow smelt  $\geq 90$  mm, and bloaters  $\geq 120$  mm were classified as "adults". Unless otherwise stated, all length measurements refer to total length.

Catches of small alewives, bloaters, and rainbow smelt are not necessarily reliable indicators of future year-class strengths for these populations, because their small size and position in the water column make them less vulnerable to bottom trawls. Nevertheless, during the bloater recovery in Lake Michigan that began in the late 1970s, our trawling survey indicated that the lake contained unusually-high abundances of age-0 bloaters, so there is some correspondence between our bottom trawl catches of age-0 prey fish and their actual abundance in the lake.

Alewife— Since its establishment in the 1950s, the alewife has become a key member of the fish community. The alewife has remained the most important constituent of salmonine diet in Lake Michigan for the last 35 years (Jude et al. 1987; Stewart and Ibarra 1991; P. Peeters, Wisconsin Department of Natural Resources, Sturgeon Bay, WI, personal communication; R. Elliott, U. S. Fish and Wildlife Service, Green Bay, WI, personal communication). Most of the alewives consumed by salmonines in Lake Michigan are eaten by chinook salmon (Madenjian et al. 2002). A commercial alewife harvest was established in Wisconsin waters of Lake Michigan in the 1960s to make use of the then extremely abundant alewife that had become a nuisance and health hazard along the lakeshore. In 1986, a quota was implemented, and as a result of these rule changes and seasonal and area restrictions, the estimated alewife harvest declined from about 7,600 metric tons in 1985 to an estimated average annual

incidental harvest of only 12 metric tons after 1990 (Mike Toney, Wisconsin Department of Natural Resources, Sturgeon Bay, personnel communication). There is presently no commercial fishery for alewives in Lake Michigan.

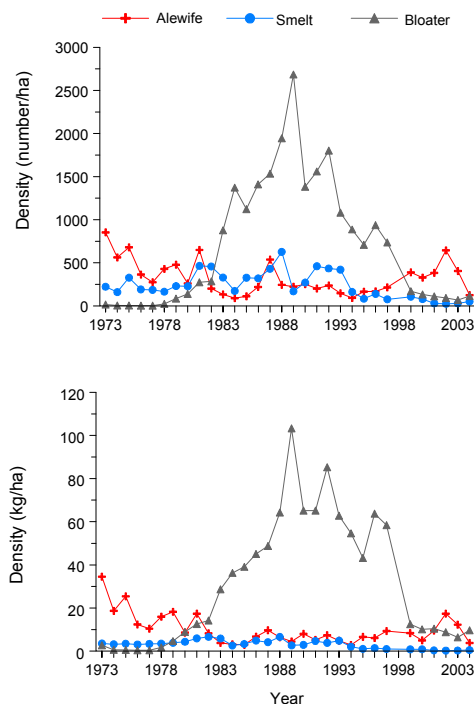


Figure 2. Density of adult alewives, rainbow smelt and bloaters as number (top) and weight (bottom) of fish per ha in Lake Michigan, 1973-2004.

Adult alewife abundance declined by nearly 70% between 2003 and 2004. Measured as number of fish per ha, density of adult alewives in Lake Michigan decreased from 403 fish per ha in 2003 to 123 fish per ha in 2004 (Figure 2; Appendix 1). Similarly, alewife density, on a weight basis, decreased from 12.3 kg per ha in 2003 to 3.9 kg per ha in 2004 (Figure 2). Only in years 1983, 1984, 1985, and 1994 was adult alewife density less than that observed in 2004. Given that predation by salmon and trout appears to be the most important factor regulating alewife abundance in Lake Michigan (Madenjian et al. 2002, 2005), a recent increase in chinook salmon biomass may be a likely cause for the recent pronounced decrease in adult alewife abundance.

RSEs for adult alewife density ranged from 10% to 42% (Figure 3). During 1973-2004, RSE averaged 21% for density on a number of fish per ha basis and 19% for density on a kg of fish per ha basis. Relatively high RSEs during 2003 and

2004 suggested that adult alewives were patchier in their spatial distribution during these last two years than during most of the previous years.

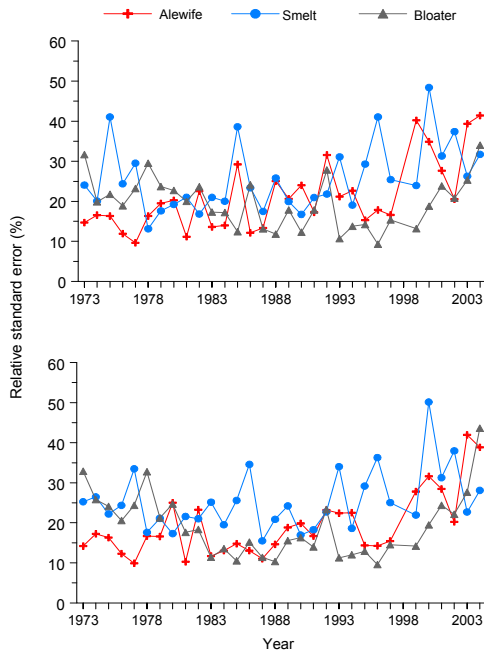


Figure 3. RSE for density of adult alewives, rainbow smelt and bloaters as number (top) and weight (bottom) of fish per ha in Lake Michigan, 1973-2004.

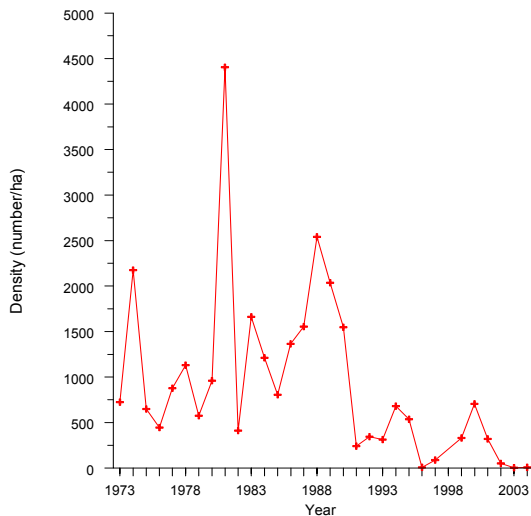


Figure 4. Density of age-0 alewives in Lake Michigan, 1973-2004.

Density of age-0 alewives, as measured by the bottom trawl survey, was relatively low at only 9 fish per ha in 2004 (Figure 4).

Again, the 1998 year-class appeared to be an exceptionally strong one, as this year-class

dominated the catch of adult alewives during 2004 (Figures 5 and 6). The 1998 year-class also dominated the catch of age-1 and older alewives during 1999-2003 (Madenjian et al. 2004). The 2002 year-class was well represented in the trawl catch during 2004 (Figures 5 and 6).

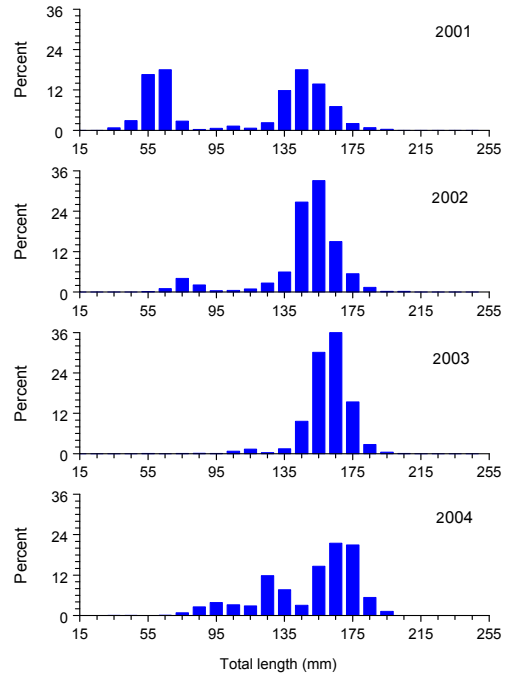


Figure 5. Length-frequency distribution of alewives caught in bottom trawls in Lake Michigan, 2001-2004.

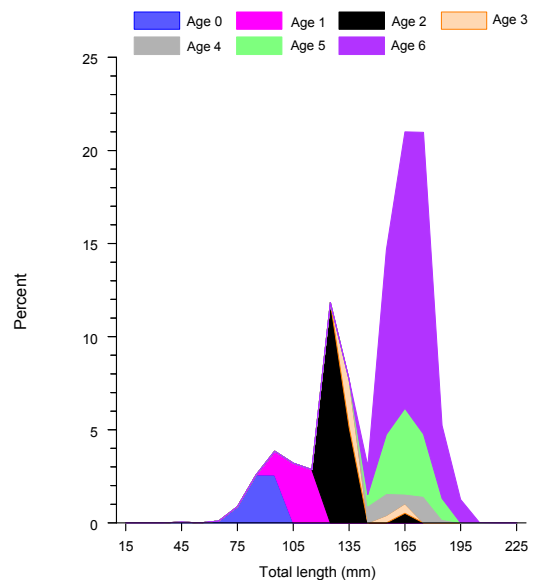


Figure 6. Age-length distribution of alewives caught in bottom trawls in Lake Michigan, 2004. The 1998 year-class is age-6 fish.

**Bloater** - Bloaters are eaten by salmonines in Lake Michigan, although not to the extent that adult alewives are consumed. Over 30% of the diet of large ( $\geq 600$  mm) lake trout at Saugatuck and on Sheboygan Reef was composed of adult bloaters during 1994-1995, although adult bloaters were a minor component of lake trout diet at Sturgeon Bay (Madenjian et al. 1998). When available, juvenile bloaters have been a substantial component of salmon and nearshore lake trout diets, particularly for intermediate-sized fish (Elliott 1993; Rybicki and Clapp 1996). The bloater population in Lake Michigan also supports a valuable commercial fishery.

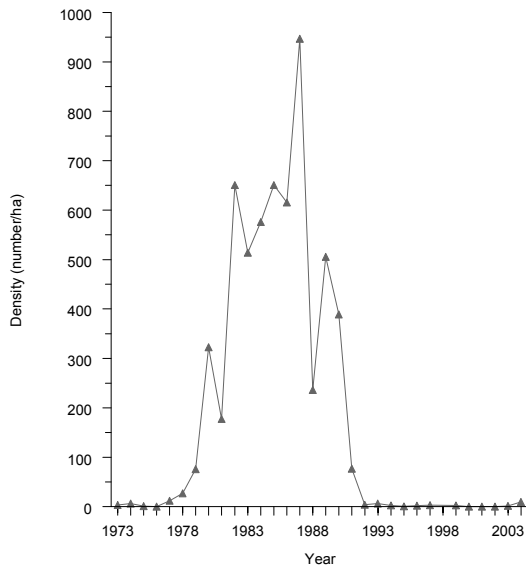


Figure 7. Density of age-0 bloaters in Lake Michigan, 1973-2004.

The overall trend in adult bloater density during 1989-2004 was a dramatic decline (Figure 2). This decline was attributable to relatively poor recruitment during 1992-2003 (Madenjian et al. 2002, 2004). Adult bloater density increased from 6.4 kg per ha in 2003 to 9.7 kg per ha in 2004 (Figure 2). Age-0 bloater density ranged from 0.2 to 6.5 fish per ha during 1992-2003, whereas age-0 bloater density ranged from 177 to 947 fish per ha between during 1980-1990 (Figure 7). In 2004, age-0 bloater density was 9.6 fish per ha, which was the highest value in the time series since 1991. Moreover, age-0 bloater density in 1977, the year in which the bloater recovery in Lake Michigan began (Eck and Wells 1987), was 11.8 fish per ha. Madenjian et al. (2002) have proposed that the Lake Michigan bloater population may be cycling in abundance, with a

period of about 30 years. If bloater abundance is truly cyclical and year 2004 signaled the initiation of a bloater recovery, then age-0 bloater density would be expected to substantially increase in 2005.

**Rainbow smelt** – Adult rainbow smelt is an important diet item for intermediate-sized (400 to 600 mm) lake trout in the nearshore waters of Lake Michigan (Stewart et al. 1983; Madenjian et al. 1998). Overall, however, rainbow smelt are not consumed by Lake Michigan salmonines to the same extent as alewives. The rainbow smelt population supports commercial fisheries operated in Wisconsin and Michigan waters (Belonger et al. 1998; P. Schneeberger, Michigan Department of Natural Resources, Marquette, MI, personal communication).

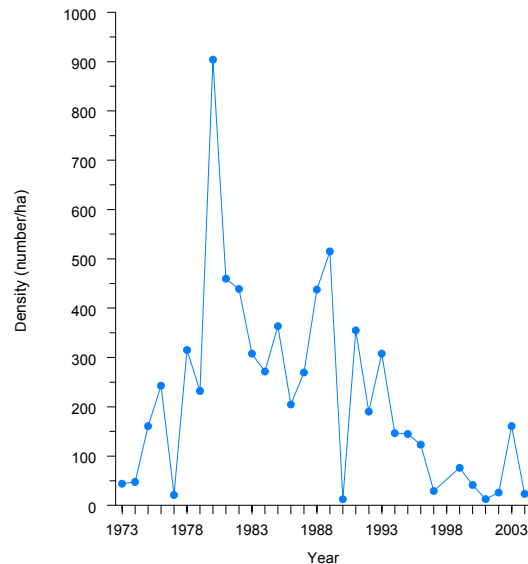


Figure 8. Density of age-0 rainbow smelt in Lake Michigan, 1973-2004.

Adult rainbow smelt density increased from 30 fish per ha in 2003 to 50 fish per ha in 2004 (Figure 2). On a weight basis, density increased from 0.3 to 0.5 kg per ha between 2003 and 2004. Adult rainbow smelt abundance declined substantially from 1992 to 1997, and has remained low since 1997 (Figure 2). Causes for the decline remain unclear. Consumption of smelt by salmonines was higher in the mid 1980s than during 1992-1997 (Madenjian et al. 2002), yet adult smelt abundance remained high during the 1980s. Average density of age-0 rainbow smelt during 1973-1979 was not appreciably higher than



average abundance of age-0 rainbow smelt during the 1990s (Figure 8). Age-0 smelt abundance substantially decreased from 161 fish per ha in 2003 to 23 fish per ha in 2004 (Figure 8). Interpretation of the long-term time series for adult rainbow smelt density remains difficult.

**Sculpins** – The cottid populations in Lake Michigan proper are dominated by deepwater, and to a lesser degree, slimy sculpins. Spoonhead sculpins, once fairly common, suffered declines to become rare to absent by the mid 1970s (Eck and Wells 1987). Spoonhead sculpins are still encountered in Lake Michigan, but in small numbers (Potter and Fleischer 1992).

Slimy sculpin is a favored prey of juvenile lake trout in nearshore regions of the lake (Stewart et al. 1983; Madenjian et al. 1998). As lake trout grow, the importance of sculpins in lake trout diet decreases substantially so that sculpins form only a minor portion of adult lake trout diet. Deepwater sculpin is an important diet item for burbot in Lake Michigan, especially in deeper waters (Van Oosten and Deason 1938; Brown and Stedman 1995; Fratt et al. 1997).

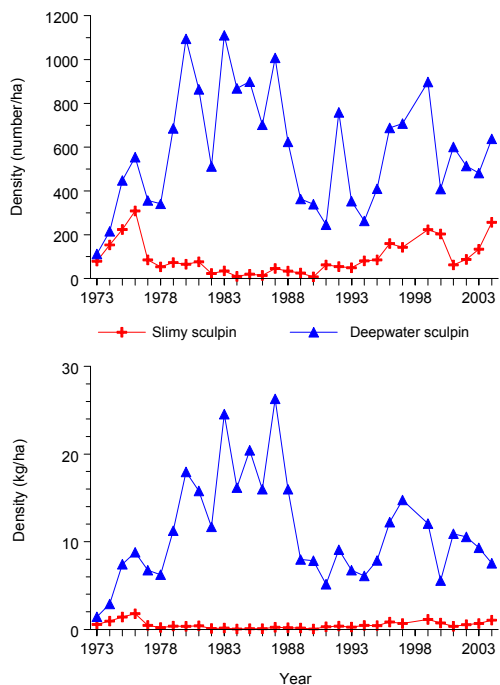


Figure 9. Density of slimy and deepwater sculpins as number (top) and weight (bottom) of fish per ha in Lake Michigan, 1973-2004.

Density of deepwater sculpins in Lake Michigan increased to 638 fish per ha in 2004, compared with 481 fish per ha in 2003 (Figure 9). On a weight basis, density of deepwater sculpins decreased from 9.3 to 7.5 kg per ha between 2003 and 2004. From a standpoint of number of fish per ha, deepwater sculpin density has increased somewhat during 1990-2004. From a standpoint of kg of fish per ha, deepwater sculpin density has just increased slightly or leveled off between 1990 and 2004 (Figure 9). Leveling off of deepwater sculpin abundance during the 1990s coincided with a leveling off of burbot abundance.

Density of slimy sculpins in Lake Michigan increased from 134 fish per ha in 2003 to 257 fish per ha in 2004 (Figure 9). Overall, slimy sculpin abundance showed an increasing trend during the 1990s. This increase in abundance may have actually begun in 1986, when an emphasis was first placed on stocking lake trout on offshore reefs rather than stocking lake trout in areas closer to shore in Lake Michigan. Slimy sculpin is a favored prey of juvenile lake trout. The GLSC bottom trawl survey does not cover the rocky, offshore reefs that have been heavily stocked with lake trout since 1986. Thus, the observed increase during the 1990s in slimy sculpin abundance detected in the GLSC bottom trawl survey was likely attributable to the emphasis on stocking lake trout on offshore reefs beginning in 1986 (Madenjian et al. 2002). Slimy sculpin abundance appeared to be leveling off during 1999-2004. *Diporeia* has dominated the diet of slimy sculpins in Lake Michigan since the 1970s (Madenjian et al. 2002), and *Diporeia* abundance in Lake Michigan has declined during the 1990s (Nalepa et al. 2004). Whether the decrease in *Diporeia* abundance will eventually have a negative impact on slimy sculpin abundance remains to be determined.

Analysis of the bottom trawl survey data has indicated that alewives interfering with deepwater sculpin reproduction and predation by burbot on deepwater sculpins have been the most important factors affecting deepwater sculpin abundance in Lake Michigan (Madenjian et al. In review). The survey data provided no evidence to support the contention that slimy sculpins exerted a significant negative effect on deepwater sculpin abundance.

## BIOMASS

We estimated a total lake-wide biomass of prey fish available to the bottom trawl in 2004 of 80.073 kilotonnes (kt) (1 kt = 1000 metric tons) (Figure 10). This total prey fish biomass was the sum of the population biomass estimates for alewife, bloater, rainbow smelt, deepwater sculpin, and slimy sculpin. Bloaters constituted 43% (34.321 kt), deepwater sculpins constituted 33% (26.519 kt), and alewives constituted 17% (13.721 kt) of the total prey fish biomass in Lake Michigan in 2004.

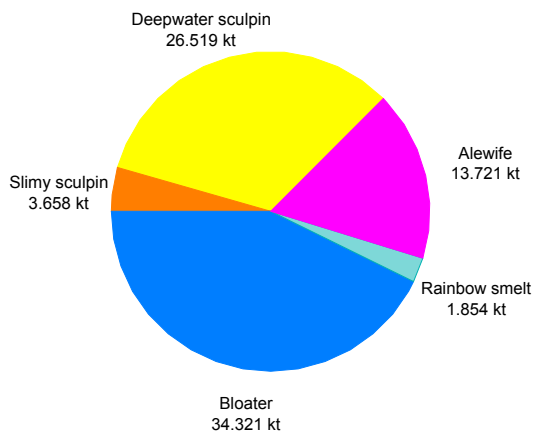


Figure 10. Estimated lake-wide biomass of prey fishes in Lake Michigan, 2004, based on bottom trawl surveys.

Total prey fish biomass in Lake Michigan has shown a declining trend since 1989, and this decline is mainly attributable to the tremendous decrease in bloater biomass (Figure 11). The current bloater biomass is about 9% of the peak value in 1989. Total prey fish biomass did increase slightly between 2000 and 2002 due to an increase in alewife biomass (Figure 11). The slight decline in total prey fish biomass between 2002 and 2004 was primarily due to a decrease in alewife biomass. The slight increase in total prey fish biomass time series during 2000-2002 was due to the exceptionally large 1998 alewife year-class. Nonetheless, alewife biomass declined during the 1970s, but fluctuated with no consistent trend during 1982-2004. Long-term trends in alewife biomass suggested that the salmonine stocking program was not only effective in

reducing alewife abundance in Lake Michigan, but also effective in maintaining relatively low alewife abundance for the last 22 years (Madenjian et al. 2002). Rainbow smelt biomass declined between 1992 and 1997, and has remained low from 1997 to 2004. Deepwater sculpin biomass has fluctuated without trend during 1990-2004 (Figure 11).

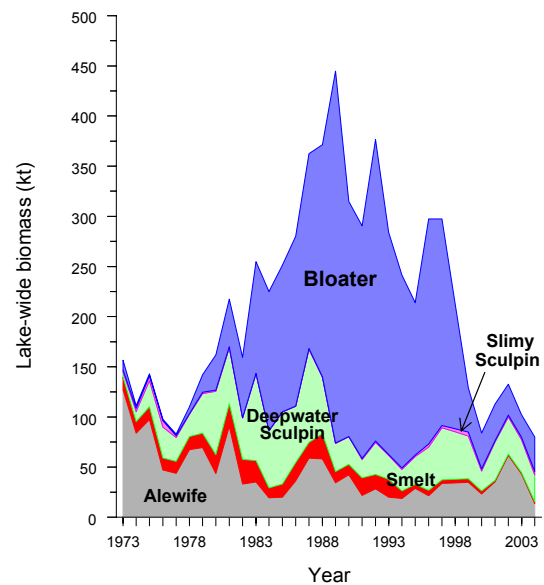


Figure 11. Estimated lake-wide biomass of prey fishes in Lake Michigan, 1973-2004, based on bottom trawl surveys.

## OTHER SPECIES OF INTEREST

**Burbot** – Burbot and lake trout represent the native top predators in Lake Michigan. The decline in burbot abundance in Lake Michigan during the 1950s has been attributed to sea lamprey predation (Wells and McLain 1973). Sea lamprey control was a necessary condition for recovery of the burbot population in Lake Michigan, however Eshenroder and Burnham-Curtis (1999) proposed that a reduction in alewife abundance was an additional prerequisite for burbot recovery.

Burbot collected in the bottom trawls are typically large individuals (>350 mm TL); juvenile burbot apparently inhabit areas not covered by the bottom trawl survey.

According to the GLSC bottom trawl survey data, the recovery of burbot in Lake Michigan appears to be complete. After a period of initial low

abundance, catches of burbot in the bottom trawls increased sharply from 1983 to 1990 (Figure 12). Burbot catch leveled off during 1990-2001, and then decreased during 2001-2003. The decrease in burbot density during 2001-2003 may have been partly due to increased predation by sea lampreys; lake-wide estimates of spawning sea lampreys in Lake Michigan tributaries increased during 2000-2003 (D. Lavis, U. S. Fish and Wildlife Service, Ludington, MI, personal communication). Burbot density increased from 0.43 fish per ha in 2003 to 0.58 fish per ha in 2004 (Figure 12).

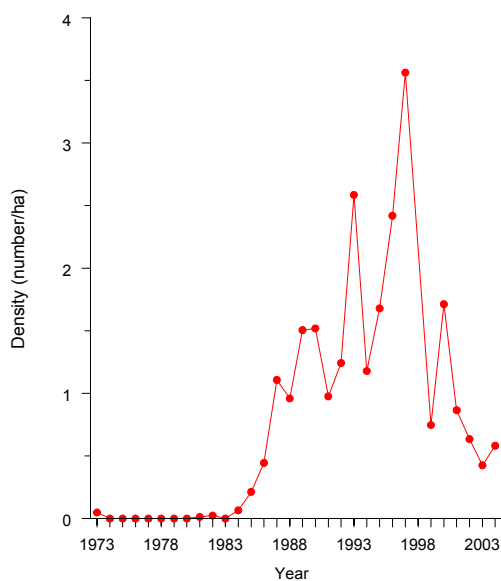


Figure 12. Density of burbot in Lake Michigan, 1973-2004.

Yellow perch – The yellow perch population in Lake Michigan has supported valuable recreational and commercial fisheries (Wells 1977). GLSC bottom trawl surveys provide an index of age-0 yellow perch abundance, which serves as an indication of yellow perch recruitment success. According to the standard trawl series, recruitment success was poor in most years during 1989-2004 (Figure 13). No age-0 yellow perch were caught in 2004. Most researchers believe that a combination of several factors are responsible for this prolonged period of low recruitment success; these factors include poor weather conditions in some years, a low abundance of female spawners in some years, and

possibly a low availability of zooplankton for yellow perch larvae in some years (Makauskas and Clapp 2000).

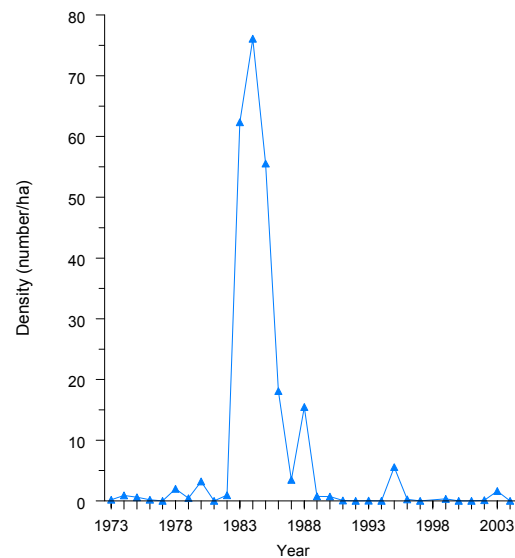


Figure 13. Density of age-0 yellow perch in Lake Michigan, 1973-2004.

Round goby – The first catch of round gobies in the GLSC bottom trawl survey of Lake Michigan occurred in 2003. In total, 23 round gobies were caught during the 2003 survey. Catches were limited to the Manistique and Saugatuck transects at depths ranging from 9 to 27 m. Round goby total lengths ranged from 57 to 161 mm. Similarly, 26 round gobies were caught during the 2004 survey, and 2004 catches were again limited to the Manistique and Saugatuck transects. Depths of capture ranged from 18 to 37 m in 2004, and round goby total lengths ranged from 58 to 161 mm.

The round goby is an invader from the Black and Caspian seas. Round gobies have been observed in bays and harbors of Lake Michigan since 1993, and round gobies have been captured by Michigan DNR personnel in the southern main basin of the lake as early as 1997 (Clapp et al. 2001). With additional years of continued surveillance, results from the GLSC bottom trawl survey should prove useful in detecting significant effects of round gobies on the Lake Michigan fish community.

Dreissenid mussels – The first zebra mussel noted in Lake Michigan was found in May 1988



(reported in March 1990) in Indiana Harbor at Gary, Indiana. By 1990, adult mussels had been found at multiple sites in the Chicago area, and by 1992 were reported to range along the eastern and western shoreline in the southern two-thirds of the lake, as well as in Green Bay and Grand Traverse Bay (Marsden 1992). In 1999, catches of dreissenid mussels in our bottom trawls became significant and we began recording weights from each tow. Lake Michigan dreissenid mussels include two species: the zebra mussel and the quagga mussel. The quagga mussel is a more recent invader to Lake Michigan than the zebra mussel (Nalepa et al. 2001). According to the GLSC bottom trawl survey, lake-wide biomass of dreissenid mussels trended neither upward nor downward during 1999-2004 (Figure 14). Estimated lake-wide biomass of dreissenid mussels increased from 13.9 kt in 2003 to 27.8 kt in 2004. The dreissenid mussel invasions have been associated with the decline in the amphipod *Diporeia* in Lake Michigan, although the mechanism by which dreissenid mussels are negatively affecting *Diporeia* remains unidentified (Madenjian et al. 2002).

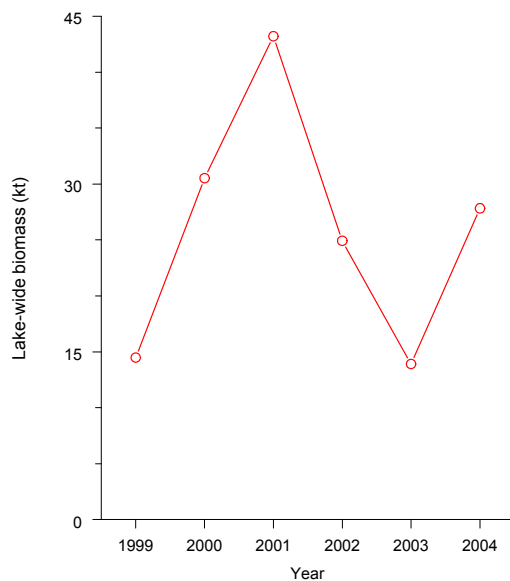


Figure 14. Estimated lake-wide biomass of dreissenid mussels in Lake Michigan, 1999-2004, based on bottom trawl surveys.

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Appendix 1. Mean density and lake-wide biomass estimates for various fishes and dreissenid mussels in Lake Michigan during 2004. Estimates based on the bottom trawl survey. Standard error enclosed in parentheses. NA denotes that estimate is not available.

Taxon	Density (number/ha)	Density (kg/ha)	Lake-wide biomass (kt)
age-0 alewife	9.18 (5.79)	0.043 (0.028)	0.152 (0.099)
adult alewife	123.43 (51.10)	3.853 (1.496)	13.568 (5.270)
age-0 bloater	9.59 (3.63)	0.079 (0.030)	0.277 (0.104)
adult bloater	114.36 (38.89)	9.667 (4.219)	34.043 (14.858)
age-0 rainbow smelt	23.47 (10.49)	0.021 (0.005)	0.074 (0.019)
adult rainbow smelt	50.44 (16.00)	0.505 (0.142)	1.780 (0.500)
deepwater sculpin	637.98 (237.41)	7.530 (2.091)	26.519 (7.364)
slimy sculpin	257.13 (64.56)	1.039 (0.307)	3.658 (1.082)
burbot	0.58 (0.25)	0.875 (0.356)	3.081 (1.255)
age-0 yellow perch	0 (0)	0.000 (0.000)	0.000 (0.000)
dreissenid mussels	NA	7.903 (1.074)	27.831 (3.781)